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1 **Integration of gene expression and DNA methylation identifies epigenetically controlled**
2 **modules related to PM_{2.5} exposure**

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102

103 **Abstract**

104 Air pollution has been associated with adverse health effects across the life-course. Although
105 underlying mechanisms are unclear, several studies suggested pollutant-induced changes in
106 transcriptomic profiles. In this meta-analysis of transcriptome-wide association studies of 656
107 children and adolescents from three European cohorts participating in the MeDALL
108 Consortium, we found two differentially expressed transcript clusters (FDR $p < 0.05$)
109 associated with exposure to particulate matter < 2.5 micrometers in diameter ($PM_{2.5}$) at birth,
110 one of them mapping to the *MIR1296* gene. Further, by integrating gene expression with
111 DNA methylation using Functional Epigenetic Modules algorithms, we identified 9 and 6
112 modules in relation to $PM_{2.5}$ exposure at birth and at current address, respectively (including
113 *NR1I2*, *MAPK6*, *TAF8* and *SCARA3*). In conclusion, $PM_{2.5}$ exposure at birth was linked to
114 differential gene expression in children and adolescents. Importantly, we identified several
115 significant interactome hotspots of gene modules of relevance for complex diseases in relation
116 to $PM_{2.5}$ exposure.

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127 **Key words:** air pollution, DNA methylation, gene expression, integration, children

128 **1. Introduction**

129 Air pollution exposure at birth has been associated with different types of health effects, such
130 as adverse pregnancy outcomes (Pedersen et al. 2013), childhood airway disease (Gehring et
131 al. 2015), and neurodevelopmental disorders (Sram et al. 2017). Although the precise
132 mechanisms responsible for these health effects are unclear, several studies have suggested
133 oxidative stress and systemic inflammation as potential intermediate biological responses to
134 air pollutants (Kelly and Fussell 2015). Accumulating evidence suggests that these acute
135 systemic effects of long-term exposure to air pollution can be detected by assessing genome-
136 wide gene expression profiles in peripheral blood cells (Mostafavi et al. 2017).

137 Exposure to air pollutants has been shown to induce changes in gene expression in animal and
138 *in vitro* experiments (Bhetraratana et al. 2019; Kim et al. 2019; Zhu et al. 2019), but evidence
139 from human studies is scarce. Short-term inhalation studies reported changes in the expression
140 of genes involved in inflammation, tissue growth and host defense against environmental
141 insults, including IGF-1 signaling, insulin receptor signaling and NRF2-mediated oxidative
142 stress response pathway in blood (Huang et al. 2010; Peretz et al. 2007), as well as genes
143 associated with bronchial immune responses in bronchoalveolar lavages, protein degradation,
144 and coagulation (e.g. *PLAU*, *F2R*, *CBL*, *UBRI*) (Pettit et al. 2012) in response to exposure to
145 diesel exhaust. A genome-wide gene expression microarray analysis of 63 non-smoking
146 employees at 10 trucking terminals in the northeastern US identified a set of genes implicated
147 in ischemic heart disease, chronic obstructive pulmonary disease (COPD), lung cancer, and
148 other pollution-related illnesses (Chu et al. 2016). Another study based on 550 healthy
149 subjects participating in cohorts from Italy and Sweden has shown differential gene
150 expression related to long-term exposure to nitrogen oxides (NO_x) similar to tobacco induced
151 changes in the transcriptome (Mostafavi et al. 2017). In a study of school age children
152 inhabiting a severely polluted area from the Czech Republic, numerous genes were found to

153 have their expression in blood relatively increased when compared with children from a rural
154 less polluted area (van Leeuwen et al. 2006). In other studies, gene-environment interactions
155 with respect to air pollution exposure and expression levels have been reported (Fave et al.
156 2018; Gref et al. 2017).

157 To our knowledge, no study has evaluated expression levels across the genome in children
158 and adolescents in relation to individual ambient air pollution exposure. Epigenomic changes
159 related to air pollution are, on the other hand, relatively well studied (Gruzieva et al. 2017;
160 Gruzieva et al. 2019). Although several differentially methylated CpGs and regions have been
161 identified in relation to environmental factors, most studies found rather weak associations
162 with ambient air pollution exposure (Gruzieva et al. 2017; Gruzieva et al. 2019; Lee et al.
163 2019; Plusquin et al. 2017). Integrating different types of omics data may shed new light on
164 gene modules or molecular pathways, which play key roles in cellular responses and
165 subsequent disease (Jiao et al. 2014; Tian et al. 2017). In addition, it has been suggested that
166 omics integration may enhance study power with increased likelihood to identify
167 biologically relevant mechanisms and similarity patterns between groups of subjects (Li et
168 al. 2018; Wang et al. 2014). For example, in a recent integrative study of methylome and
169 transcriptome of human cardiomyocytes, multiple altered methylome and transcriptome
170 signatures in the cardiac disease-specific genes, following exposure to particulate matter
171 less than 2.5 micrometers in diameter (PM_{2.5}), have been reported (Yang et al. 2018).

172 Human studies of environmental exposures combining different -omics data are, however,
173 scarce (Vargas et al. 2018).

174 In the present study, we explored the association of gene expression profiles with PM_{2.5}
175 exposure at birth and at the time of blood sampling (i.e. pre-school age or adolescence).
176 Moreover, we performed an integrative analysis of gene expression and DNA methylation
177 data to identify molecular pathways that are epigenetically and functionally affected by PM_{2.5}

178 exposure. Within the framework of the European collaborative Mechanisms of the
179 Development of ALLergy (MeDALL) project (Bousquet et al. 2011), we used harmonized
180 data of three European birth cohort studies for which a standardized assessment of air
181 pollution exposure for PM_{2.5} was available.

182

183 **2. Methods**

184 **2.1 Study population**

185 Data from three European birth cohort studies from Sweden (BAMSE (Schultz et al. 2016));
186 Germany (GINIplus (Heinrich et al. 2017)); and Spain (INMA (Guxens et al. 2011)),
187 participating in the MeDALL project (Bousquet et al. 2011) were included in the present
188 analysis (n total = 656). All cohorts acquired ethics approval and informed consent from
189 participants prior to data collection through local ethics committees.

190

191 **2.2 Air pollution exposure assessment**

192 In the MeDALL cohorts, annual average concentrations of PM_{2.5} were estimated at home
193 addresses at birth and at the time of bio-sampling (i.e. pre-school age in INMA and
194 adolescence in BAMSE and GINIplus) through land-use regression (LUR) models developed
195 for each study area within the European Study of Cohorts for Air Pollution Effects (ESCAPE)
196 project (Eeftens et al. 2012). Further details about exposure assessment are provided in the
197 online supplement.

198

199 **2.3 Gene expression measurements**

200 Expression levels were measured in BAMSE (244 individuals of mean age 16.69 years),
201 GINIplus Munich (247 individuals aged 15.2), and INMA (165 individuals aged 4.48) cohorts
202 with the Affymetrix Human Transcriptome Array 2.0 (HTA) (Lemonnier et al. 2020). Whole

203 blood was collected from study subjects in PAXgene tubes, and RNA was extracted batch-
204 wise (QIAGEN, Courtaboeuf, France). RNA yield and quality were assessed with state-of-
205 the-art spectrophotometry and lab-on-a-chip microfluidic technologies, respectively. RNA of
206 highest quality was selected for amplification, labelling, and hybridization on HTA with WT
207 PLUS kit (Affymetrix Inc.) at the CNRS USR 3010 unit in Lyon. Gene expression levels were
208 normalized with the Robust Multi-array Average (RMA) algorithm including quantile
209 normalization (Irizarry et al. 2003), and version 36 of NetAffx annotation was used to
210 annotate the 67,528 transcript clusters (probes covering a region of the genome reflecting all
211 the exonic transcription evidence known for the region and corresponding to a known or
212 putative gene, including coding and non-coding genes). The empirical Bayes method was
213 applied for batch correction on the main three production phases using ComBat from the sva
214 package in R (Johnson et al. 2007). More detailed information about the gene expression
215 sample assessment can be found elsewhere (Lemonnier et al. 2020). The estimated cell type
216 proportions were calculated from gene expression data using CIBERSORT
217 (<http://cibersort.stanford.edu/>) (Newman et al. 2015).

218

219 **2.4 DNA methylation measurements**

220 Epigenome-wide DNA methylation was measured using DNA extracted from whole blood
221 samples collected at the age of 16 years in the BAMSE (n=262) and at 4 years in the INMA
222 (n=201) cohorts (Xu et al. 2018) processed with the Illumina Infinium HumanMethylation450
223 BeadChip (Illumina Inc., San Diego, USA). More information is provided in the online
224 supplement.

225

226 **2.5 Statistical analyses**

227 *2.5.1 Genome-wide gene expression analyses*

228 We examined the associations between exposure to air pollutants and gene expression levels
229 across the genome by means of linear regression analysis using the limma package in R
230 (Ritchie et al. 2015). Air pollution concentrations were entered as continuous variables
231 without transformation. All analyses were adjusted for sex, age (years), study region (where
232 applicable), maternal education, maternal smoking during pregnancy, second-hand smoking,
233 active smoking status (in adolescence), BMI (kg/m²), physical activity at the time of bio-
234 sampling, season of blood sampling and cell type. Cohort-specific results of the TWAS were
235 subsequently included in a fixed-effect meta-analysis by combining study-specific weights
236 based on the inverse of the variance. We adjusted for multiple testing using the Benjamini &
237 Hochberg false discovery rate (FDR) correction (Benjamini and Hochberg 1995). The results
238 below FDR P<0.05 threshold were labeled as statistically significant. The results are
239 presented per 5 µg/m³ increase in PM_{2.5}. A pearson correlation analysis was performed
240 between PM_{2.5} at birth and at the time of biosampling for each cohort. All cohort-specific
241 statistical analyses were performed using R (R Development Core Team 2019) and
242 Bioconductor packages (Gentleman et al. 2004), and the meta-analysis was performed using
243 METAL software (Willer et al. 2010).

244

245 *2.5.2 Integration of DNA methylation and gene expression results*

246 An integration analysis of genome-wide DNA methylation with matched gene expression data
247 available in BAMSE and INMA was performed on 240 and 103 subjects, respectively. First,
248 we combined cohort-specific associations of air pollution exposure with genome-wide DNA
249 methylation in the BAMSE and INMA cohorts (no DNA methylation data were available in
250 GINIplus) after adjustment for potential covariates and cell proportions, in a fixed-effect
251 meta-analysis. In the case of Illumina 450k data, we assigned to a gene the average value of
252 probes mapping to within 200bp of the TSS. If no probes mapped to within 200bp of the TSS,

253 we used the average of probes mapping to the 1st exon of the gene. If such probes were not
254 present, we used the average of probes mapping to within 1500bp of the TSS. Transcript
255 clusters were mapped to genes as described above. The integration of the genes identified in
256 the TWAS and EWAS meta-analysis, along with protein–protein interaction (PPI) was
257 performed using the Functional Epigenetic Module (FEM) algorithm in R package (Jiao et al.
258 2014). The FEM algorithm first construct an integrated network with weights on the network
259 edges from the associations between PM_{2.5} and both gene expression and DNA methylation,
260 and then inference of the FEMs as heavy subgraphs on this weighted network (Jiao et al.
261 2014). The integration of DNA methylation and gene expression profiling was performed
262 after constructing the PPI network for hub gene identification from the High-quality
263 INTeractomes (HINT) database (<http://hint.yulab.org/>) (Das and Yu 2012). HINT is a
264 database of high-quality PPIs from 8 interactome resources (BioGRID, MINT, iRefWeb, DIP,
265 IntAct, HPRD, MIPS and the PDB) consisting of two types of interactions: binary physical
266 interactions and co-complex associations from different organisms. In total, 150,199
267 interactions were obtained after removing duplicates and self-linked interactions.

268

269 *2.5.3 Functional analyses*

270 Kyoto Encyclopaedia of Genes and Genomes (KEGG) (Kanehisa et al. 2019) pathway
271 enrichment analysis was performed using network-based pathway annotation tool BinoX
272 (Ogris et al. 2017) in PathwAX II web server (<http://pathwax.sbc.su.se/>) (Ogris et al. 2016).
273 The algorithm assesses the statistical significance of ‘pathway gene-set’ enrichment by
274 evaluating the amount of interactions between genes within a genome wide functional
275 association network. P-values for enrichment were adjusted for multiple testing using the
276 FDR method. Pathways with an estimated FDR P-value below 0.05 were selected as
277 significantly enriched. In addition, to understand the functional involvement of PM_{2.5}

278 exposure in gene expression we performed the KEGG pathway analyses by applying a less
279 stringent cut off P-value < 0.05 and $|\logFC| > 0.15$.

280

281 *2.5.4 Sensitivity analyses*

282 To address potential age-specific effects, we performed a sensitivity analysis limiting our
283 transcriptome-wide meta-analysis to two adolescent cohorts. Further, to rule out residual
284 confounding effects from active smoking, we performed a sensitivity analysis in the two
285 adolescent cohorts excluding active smokers and compared those with the results based on the
286 whole sample.

287

288 **3. Results**

289 In total, 656 pre-school children (INMA) and adolescents (GINIplus and BAMSE) were
290 included in the meta-analysis of genome-wide gene expression in relation to birth and current
291 $PM_{2.5}$ exposures. The characteristics of the study subjects are presented in Table 1.

292 Noteworthy, in the subset of children with gene expression measurements in the INMA
293 cohort, the proportion of children exposed to tobacco smoke prenatally, as well as at the time
294 of blood sampling, was higher compared to the other cohorts. Exposure levels, illustrated by
295 box plots in Fig.1, were on average lowest for BAMSE (mean $PM_{2.5}$ exposure at birth and at
296 current address: 7.6 and $7.3 \mu g/m^3$, respectively), and highest for INMA (14.6 and $14.0 \mu g/m^3$,
297 respectively). The correlation between birth and current exposure levels was $r=0.45$ ($p=8.46E-$
298 14), $r=0.32$ ($p=9.08E-08$), $r=0.46$ ($p=2.01E-10$) in BAMSE, GINIplus and INMA,
299 respectively.

300

301

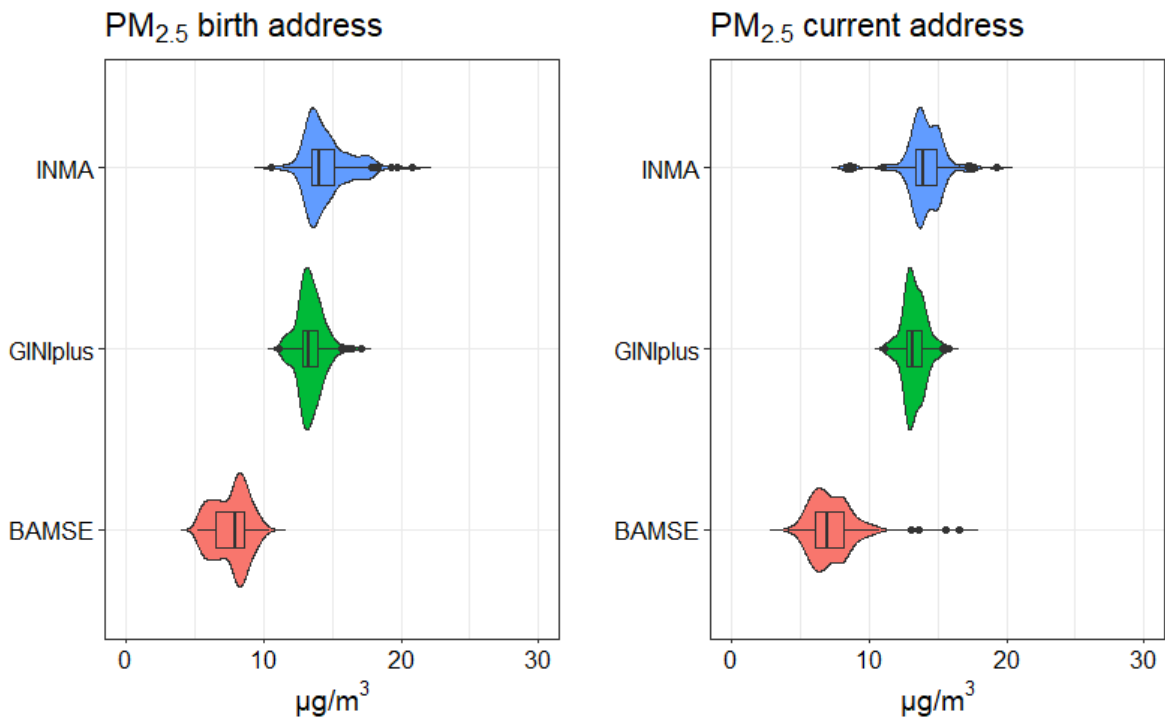
302

303 **Table 1. Characteristics of the study participants from the included cohorts.**

Characteristics	BAMSE (n=244)	GINIplus (n=247)	INMA (n=165)
Males, N (%)	136 (55.7%)	120 (48.6%)	87 (52.7%)
Age at the follow-up, yrs mean (SD)	16.7 (0.3)	15.2 (0.2)	4.5 (0.2)
Study region, N (%):	Stockholm 56 (22.9%) Järfälla 92 (37.7%) Solna 63 (25.8%) Sundbyberg 33 (13.5%)	Munich (100%)	Gipuzkoa 57 (34.5%) Sabadell 108 (65.6%)
Maternal education, ^a N (%):			
Low	79 (32.4%)	14 (5.7%)	39 (23.6%)
Medium	73 (29.9%)	87 (35.2%)	61 (36.9%)
High	92 (37.7%)	146 (59.1%)	65 (39.4%)
Season of blood sampling, N (%):			
Winter	48 (19.7%)	32 (12.7%)	28 (17.0%)
Spring	86 (35.3%)	55 (22.3%)	61 (37.0%)
Summer	42 (17.2%)	79 (32.0%)	38 (23.0%)
Autumn	68 (27.9%)	81 (33.0%)	38 (23.0%)
Physical activity, N (%):			
< 5 hours per week	120 (49.2%)	137 (55.5%)	74 (44.85%)
≥ 5 hours per week	124 (50.8%)	110 (44.5%)	91 (55.15%)
Maternal smoking during pregnancy, N (%)	24 (9.8%)	23 (9.3%)	46 (27.9%)
Second-hand smoking at the time of blood sampling, N (%)	31 (12.7%)	26 (10.5%)	84 (50.9%)
Active smoking, N (%)	25 (10.2%)	20 (8.1%)	N/A

304 N/A=Not applicable

305 ^a Low (Elementary school or 2-year high school), Medium (3-year High school) and High
306 (University).



307

308 **Fig. 1. Exposure for particulate matter (PM_{2.5}) at the birth and at the current address at**
 309 **the time of blood sampling, µg/m³.** There are two symmetrical density plots for each box
 310 plot. This is an overlay of a violin plot and a box plot.

311

312 We found genome-wide significant association (FDR $p < 0.05$) between PM_{2.5} exposure at birth

313 and gene expression for two transcript clusters, namely TC10001332.hg.1 annotated to

314 *MIR1296* gene, and TC14001976.hg.1 that is a long non-coding RNA located near *FOXA1-2*

315 (chr14:38066368-38067552) (Table 2 and forest plot in Fig. S1 and the meta-analysis for all

316 transcript clusters are presented in Supplementary Table 1). No association between current

317 PM_{2.5} exposures and differentially expressed genes (DEGs) at FDR $p < 0.05$ significance level

318 was found (top ten significant transcript clusters are presented in Supplementary Table 2).

319 However, we found that 18 transcript clusters among the top 100 significant DEGs related to

320 birth PM_{2.5} exposure, including *KRBA2*, *NRG1*, *SCAND1* and *ZNF605* genes, were also

321 significantly associated, at the nominal level ($p < 0.05$) and with the same direction, with

322 current PM_{2.5} exposure (Supplementary Table 3). The results of sensitivity analyses limited to

323 two adolescent cohorts demonstrated good agreement with those based on all three cohorts

324 (genome-wide correlation of beta coefficients = 0.69 and 0.77 in the analysis with PM_{2.5}
 325 exposure at birth and at the time of biosampling, respectively (Supplementary Fig. S2).

326 Further, we found high consistency in the results of analyses including and excluding active
 327 smokers, with a very high correlation of beta coefficients, namely 93% and 84% for PM_{2.5}
 328 exposure at birth and at the time of biosampling, respectively (Supplementary Fig. S3).

329 **Table 2. FDR-significant DEGs from the meta-analysis of the association between PM_{2.5}**
 330 **exposure at birth and gene expression in children and adolescents (n=656 from the**
 331 **BAMSE, GINIplus, and INMA cohorts).**

Probe.Set.ID	Gene	Log2FC	SE	P-value	FDR	Direction*	Het. P-value
TC10001332.hg.1	<i>MIR1296</i>	-0.19	0.04	1.74E-07	0.01	---	0.196
TC14001976.hg.1		0.42	0.09	1.02E-06	0.03	+++	0.844

332 Log2FC = The logarithm fold change (one unit of the logFC translates to a two-fold change in
 333 expression), SE = standard error, FDR = false discovery rate and Het.P-value = Heterogeneity P-value
 334 based on Cochran's Q-test for heterogeneity. Results are presented per 5 µg/m³ increase in PM_{2.5}.

335 * Order of included cohorts in the meta-analysis: BAMSE, GINIplus and INMA

336

337 As many as 364 transcript clusters mapping to 102 genes were selected from the analyses of
 338 gene expression and PM_{2.5} exposure at birth cut off P-value < 0.05 and |logFC| > 0.15. The
 339 KEGG pathways analyses based on this selection identified a few enriched pathways related
 340 to olfactory transduction, ribosome, compliment and coagulation cascades and systemic lupus
 341 erythematosus (Supplementary Table 4). From the results of analyses based on PM_{2.5}
 342 exposure at the time of biosampling, we identified 66 transcripts annotated to 13 genes
 343 involved in the ribosome-related pathway (Supplementary Table 5). Thus, the ribosome
 344 pathways were common in both PM_{2.5} exposure at birth and at the time of biosampling.
 345 Next, we used two different omics datasets (from BAMSE and INMA) to perform an
 346 integration of the genes in the transcriptome-wide analyses (TWAS) and Epigenome-Wide
 347 Association Study (EWAS) meta-analysis, along with protein–protein interaction predictions

348 from the High-quality INTeractomes (HINT) database (<http://hint.yulab.org/>) using the
349 Functional Epigenetic Module algorithms (FEMs). In these analyses we were able to identify
350 9 and 6 FEMs in relation to PM_{2.5} exposure both at birth and at the current address,
351 respectively, passing the FDR significance threshold of 0.05 (Table 3).

352

353

354 **Table 3. Summary of the output of the Functional Epigenetics Modules (FEM)**
 355 **algorithm listing 9 and 6 significant hotspots of differential methylation and expression**
 356 **in relation to birth and current PM_{2.5} exposure.**

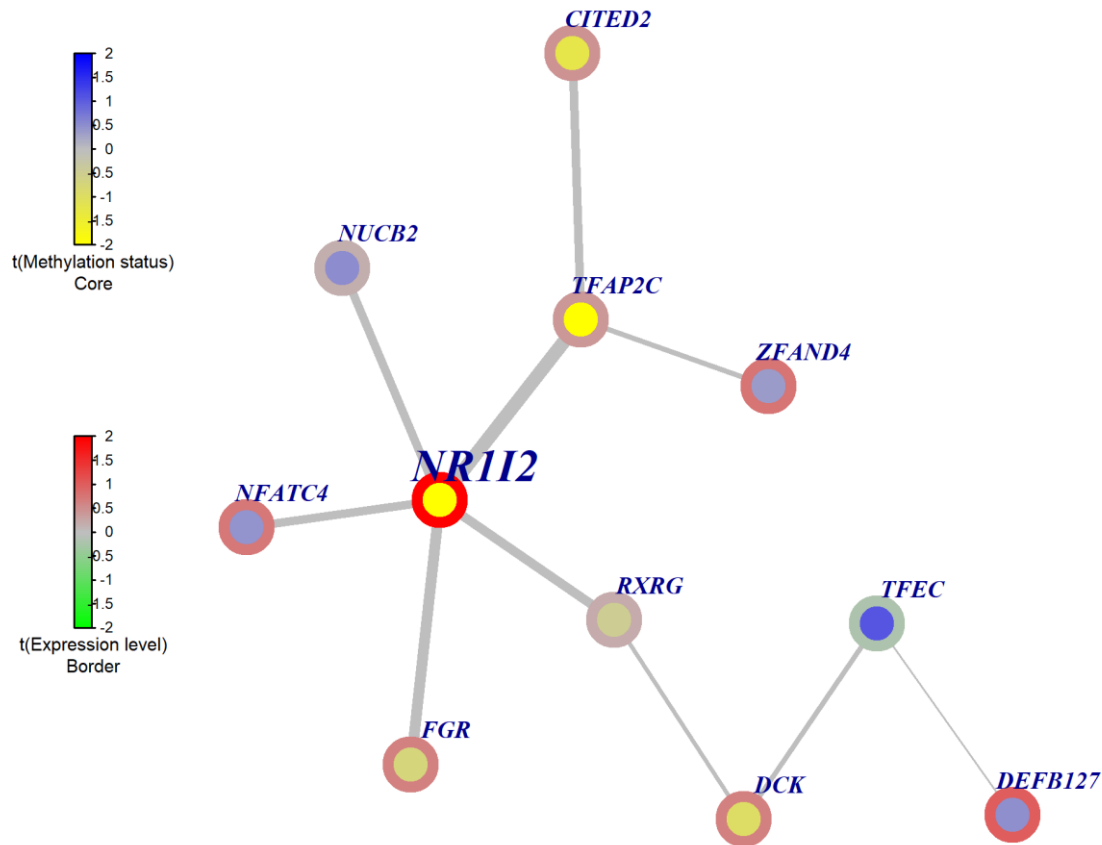
Exposure to PM _{2.5}	Seed	Size (number of genes)	Modularity	FDR *
Birth	<i>NR1I2</i>	11	2.12	0.001
	<i>SH3GL2</i>	26	2.00	0.005
	<i>TENT5A</i>	24	1.94	0.006
	<i>MAPK6</i>	64	1.95	0.010
	<i>UBE2W</i>	61	1.35	0.012
	<i>KCTD15</i>	10	1.86	0.014
	<i>MLST8</i>	10	2.11	0.018
	<i>RPP40</i>	12	1.61	0.024
	<i>GGA1</i>	14	1.59	0.025
Current	<i>TAF8</i>	20	2.17	<0.0001
	<i>TAF5</i>	20	2.17	0.001
	<i>GNAI3</i>	31	2.07	0.002
	<i>ISLR</i>	22	2.05	0.024
	<i>TRIM69</i>	10	1.84	0.032
	<i>SCARA3</i>	19	1.84	0.040

357 Columns label the seed gene symbol, the size of the FEM, its modularity (defined as the
 358 average of the edge-weights), the associated P-value. *Significant FDR p-value < 0.05.

359

360 The size of the significant modules ranged between 10 to 64 and 10 to 31 genes in the
 361 analyses with PM_{2.5} exposure at birth and current address, respectively (Supplementary
 362 Tables 6 and 7). For the top significant gene module centered around Nuclear Receptor
 363 Subfamily 1 Group I Member 2 (*NR1I2*) related to PM_{2.5} exposure at birth, we observed
 364 simultaneous hypomethylation and overexpression (Fig. 2). The KEGG pathways involving
 365 the 11 genes in *NR1I2* gene module were related to fluid shear stress and atherosclerosis,

366 protein processing in endoplasmic reticulum, adherens junction and cancer (Supplementary
367 Table 8).



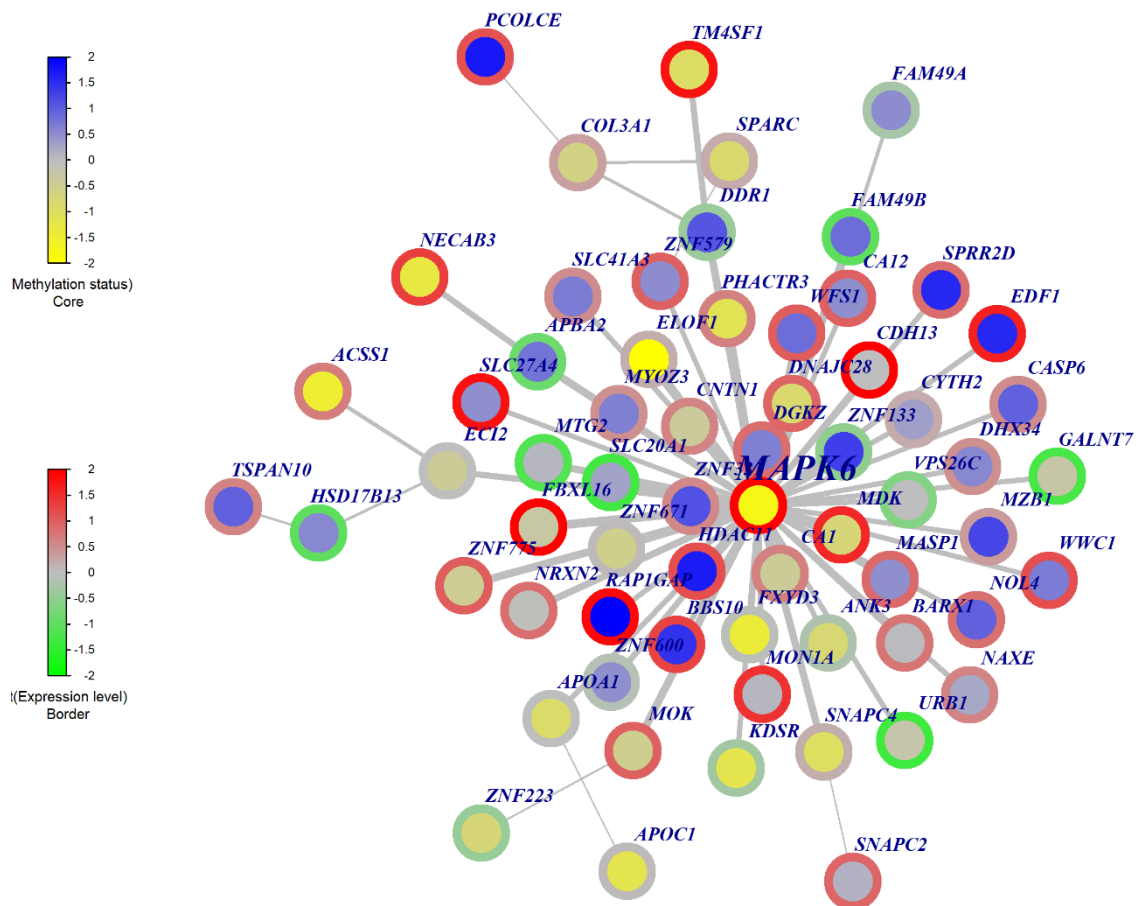
368

369 **Fig. 2. Top statistically significant functional epigenetic module around seed gene *NR112***
370 **related to PM_{2.5} exposure at birth address.**

371 The color of node represents the DNA methylation status. Blue represents hypermethylation, whereas
372 yellow represents hypomethylation. The different color of border refers to the expression patterns. The
373 red color of border represents that the genes were upregulated, whereas the green color of border
374 represents that the genes were downregulated. Edge widths are proportional to the average statistic of
375 the gene-gene interaction network.

376

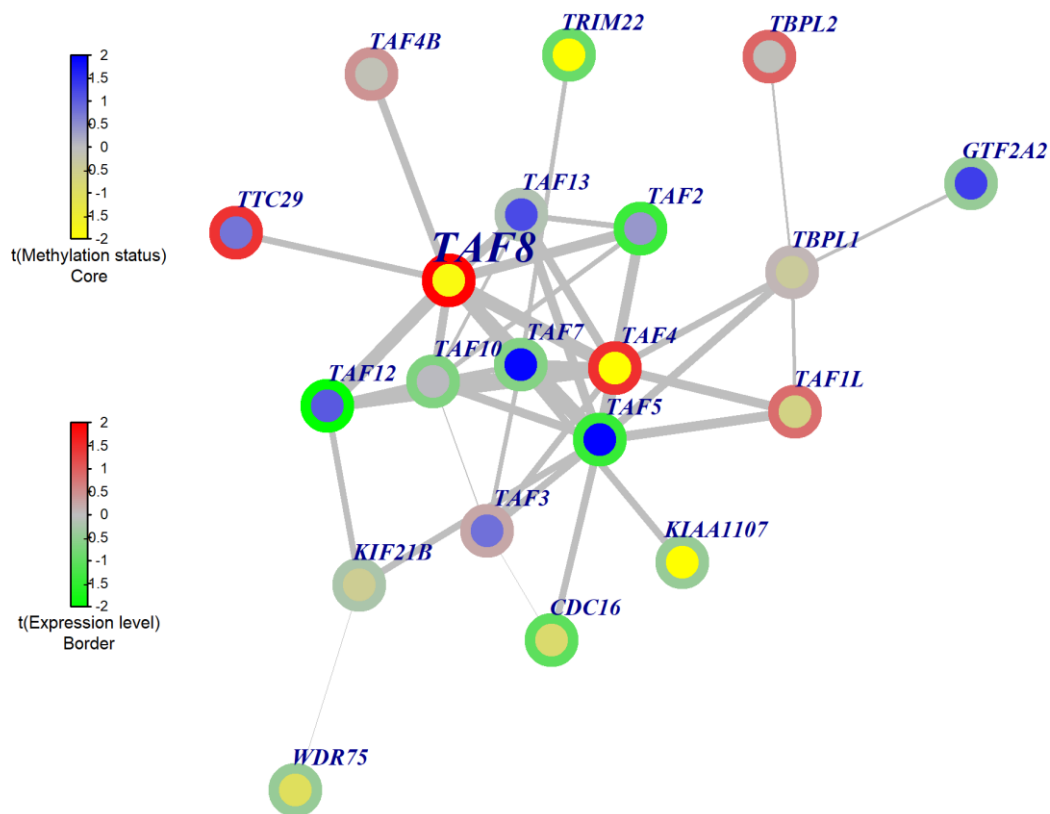
377 Another module associated with PM_{2.5} exposure at birth, mitogen-activated protein kinase 6
 378 (*MAPK6*), had the largest size of subnetwork genes (n=64) with modularity of 1.95 (Fig. 3).
 379 Subsequent KEGG pathway analysis showed that *MAPK6* module subnetwork genes are
 380 implicated in complement and coagulation cascades, as well as fatty acid degradation,
 381 oxidative phosphorylation, Huntington, Alzheimer and Parkinson disease (Supplementary
 382 Table 9).



383
 384 **Fig. 3. Top statistically significant functional epigenetic module around seed gene**
 385 ***MAPK6* related to PM_{2.5} exposure at current address.**

386 The color of node represents the DNA methylation status. Blue represents hypermethylation,
 387 whereas yellow represents hypomethylation. The different color of border refers to the expression
 388 patterns. The red color of border represents that the genes were upregulated, whereas the green color
 389 of border represents that the genes were downregulated. Edge widths are proportional to the average
 390 statistic of the gene-gene interaction network.

391 Six FEMs were identified in association with current PM_{2.5} exposure, all distinct from those
 392 found with PM_{2.5} exposure at birth. Interestingly, two highly significant gene hubs, TATA-
 393 Box Binding Protein Associated Factor 8 (*TAF8*) and TATA-Box Binding Protein Associated
 394 Factor 5 (*TAF5*) identified in relation to current PM_{2.5} exposure, contain the same 20 genes in
 395 their hubs. The *TAF8* module is shown in Fig. 4. Functional enrichment analyses of the *TAF5*
 396 and *TAF8* module subnetwork genes showed that these genes are mainly involved in basal
 397 transcription factor, cell cycle, thyroid hormones signaling pathway and notch signaling
 398 KEGG pathways (Supplementary Table 10).



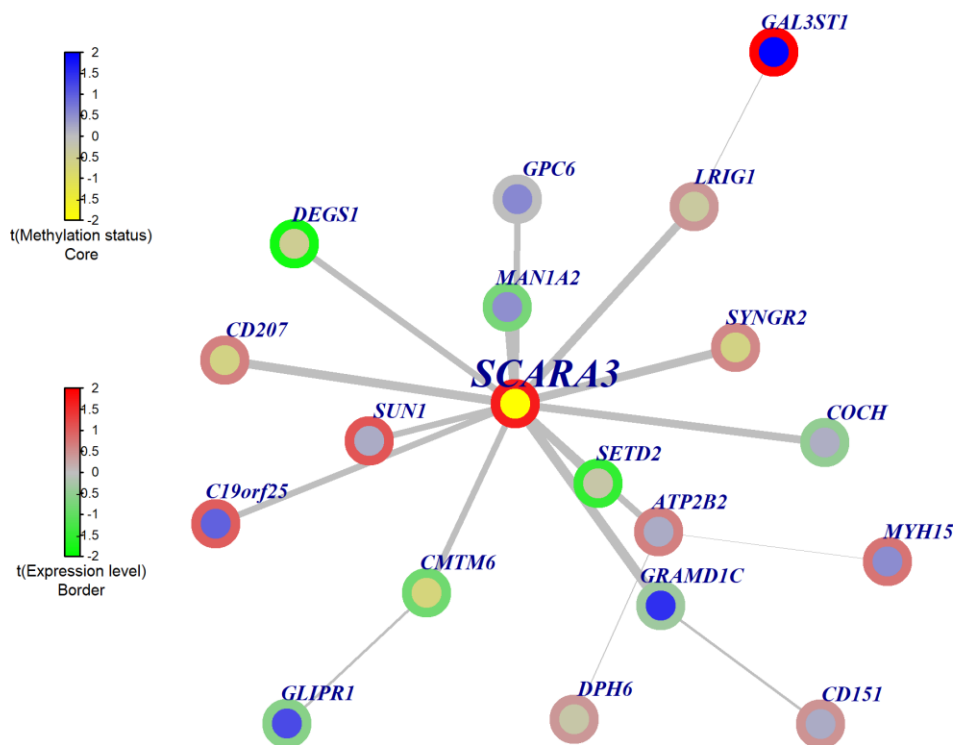
399
 400 **Fig. 4. Top statistically significant functional epigenetic module around seed gene *TAF8***
 401 **related to PM_{2.5} exposure at current address.**

402 The color of node represents the DNA methylation status. Blue represents hypermethylation, whereas
 403 yellow represents hypomethylation. The different color of border refers to the expression patterns. The
 404 red color of border represents that the genes were upregulated, whereas the green color of border

405 represents that the genes were downregulated. Edge widths are proportional to the average statistic of
406 the gene-gene interaction network.

407

408 The largest gene module size in relation to current PM_{2.5} exposure was found for *GNAI3*
409 module containing 31 genes with modularity of 2.07 in the subnetwork that appeared to be
410 involved in cancer, prion diseases, rheumatoid arthritis, tight junction, MAPK and other
411 signaling and metabolism KEGG pathways (Supplementary Table 11). Another identified
412 gene module hub associated to the current PM_{2.5} exposure, Scavenger Receptor Class A
413 Member 3 (*SCARA3*), demonstrated hypomethylation and overexpression (Fig. 5).



414

415 **Fig. 5. Statistically significant functional epigenetic module centred around seed gene**
416 ***SCARA3* related to current PM_{2.5} exposure.**

417 The color of node represents the DNA methylation status. Blue represents hypermethylation, whereas
418 yellow represents hypomethylation. The different color of border refers to the expression patterns. The
419 red color of border represents that the genes were upregulated, whereas the green color of border
420 represents that the genes were downregulated. Edge widths are proportional to the average statistic of
421 the gene-gene interaction network.

422 **4. Discussion**

423 This study represents a large-scale transcriptome-wide meta-analysis evaluating the
424 association between birth and current air pollution exposure and gene expression in children
425 and adolescents. Our meta-analysis results show associations of PM_{2.5} exposure at birth with
426 differential expression of two transcript clusters, TC10001332.hg.1, annotated to *MIR1296*
427 gene and TC14001976.hg.1, long non-coding RNA located close to *FOXAI-2* gene. However,
428 for current PM_{2.5} exposure no FDR significant DEG was found. Further, by integrating gene
429 expression and DNA methylation data into a putative protein interaction network we
430 identified several hubs linked to birth and current PM_{2.5} exposures (e.g. *NR1I2*, *MAPK6*,
431 *TAF8* and *SCARA3*).

432 To date, little is known about air pollution associated transcriptomic signatures in children.
433 Two earlier studies have investigated differential gene expression in the blood of children
434 living in areas of the Czech Republic with differing levels of air pollution. Van Leeuwen and
435 co-authors found increased blood expression of numerous genes among children inhabiting a
436 severely polluted area compared to those residing in a less polluted region (van Leeuwen et al.
437 2006). In contrast, no clear effect of exposure to air pollutants was found in a study of
438 newborns from the same geographical area (Honkova et al. 2018). These studies did not
439 assess exposure to air pollution at the individual address level, which may have resulted in
440 low statistical power. In the ENVIRONAGE birth cohort, genome-wide gene expression
441 analysis in cord blood identified fifteen transcriptomic pathways altered by prenatal PM_{2.5}
442 exposure (Winckelmans et al. 2017), including “protein processing in endoplasmic reticulum”
443 that was also observed in the present study.

444 Our transcriptome-wide meta-analysis revealed two DEGs in relation to PM_{2.5} exposure at
445 birth, representing novel associations in the context of air pollution exposure. One of them
446 *MiR-1296*, coding for miRNA that has previously been found to be linked to different types of

447 cancer, e.g. breast, hepatocellular, colorectal, prostate and lung cancer (Deng et al. 2020;
448 Majid et al. 2010; Phan et al. 2016; Tao et al. 2018; Xu et al. 2017). Also, it has been
449 proposed as a putative circulating prognostic marker of heart failure (Cakmak et al. 2015).
450 *FOXA1/2* (forkhead box a1/2) plays a key role in lung alveolar and respiratory endoderm
451 morphogenesis and differentiation, including α -cells in the endocrine pancreas, liver and
452 prostate luminal ductal epithelia (Friedman and Kaestner 2006; Lee et al. 2005; Wan et al.
453 2005; Wan et al. 2004). When comparing birth with current PM_{2.5} exposure, we found 18
454 significant DEGs related to birth PM_{2.5} exposure that were also significant at the nominal level
455 with the same direction of effect in the analysis with current PM_{2.5} exposure that includes
456 *KRBA2*, *NRG1*, *SCAND1* and *ZNF605* genes. Molecular lesions of neuregulin 1 (*NRG1*) gene
457 have been proposed as a new molecular signature of invasive mucinous adenocarcinoma of
458 the lung (Trombetta et al. 2017). It has also been repeatedly shown that PM_{2.5} exposure is
459 linked to lung cancer, particularly adenocarcinoma subtype (Raaschou-Nielsen et al. 2013).
460 Another gene, SCAN-domain-containing protein 1 (*SCAND1*) has been suggested to be
461 involved in regulation of lipid metabolism (Babb and Bowen 2003). It has also been shown
462 that the anti-oxidant and anti-inflammatory capacity of high-density lipoproteins can be
463 effected by PM_{2.5} exposures (Ramanathan et al. 2016).
464 Exploring the functional molecular patterns is critical for understanding the mechanisms of
465 biological responses to air pollution exposure. In this study, we have utilized simultaneously
466 comprehensive multi-omics profiling of biological samples, including genome-wide gene
467 expression and DNA methylation. By integrating gene expression and DNA methylation into
468 a protein interaction network we found 9 significant functional epigenetically deregulated
469 modules associated with PM_{2.5} exposure at birth. The top significant one was centered around
470 seed gene *NR1I2*, also known as steroid and xenobiotic receptor, a transcriptional regulator of
471 a number of important drug metabolizing enzymes and transporters, recently reported among

472 potential targets for asthma therapy (Wang et al. 2020). The *MAPK6* module associated with
473 $PM_{2.5}$ exposure at birth in the present study, has previously been implicated in air pollution
474 response. *In vivo*, a short-term exposure to black carbon was highly associated with epigenetic
475 changes in the promoter areas of 84 MAPK pathway genes, including *MAPK6* gene (Carmona
476 et al. 2014). Further, in a recent controlled crossover study comparing the whole blood
477 transcriptome profiles from healthy volunteers, obtained pre and post short-term exposure, to
478 high and low levels of air pollution, the authors found that activation of *MAPK6* was strongly
479 related air pollution exposure (Vargas et al. 2018).

480 Among other identified modules, we found genes also known to be associated with diseases
481 previously linked to air pollution exposure including obesity and type 2 diabetes (i.e.,
482 *KCTD15* (Lv et al. 2015; Ng et al. 2010), *MLST8* (Li et al. 2016)), as well as cardio-vascular
483 diseases (*TRIM69* (Andersson et al. 2019)). Recently, an integrative study of DNA
484 methylation and gene expression reported *SH3GL2* as one of the key genes in pathogenesis of
485 lung adenocarcinoma with hypermethylation and under expression (Jin et al. 2016) opposite
486 to our finding of $PM_{2.5}$ exposure effect of hypomethylation and overexpression. In other
487 studies, hypomethylation in certain genes, e.g. *AHRR* and *F2RL3*, has been reported as a
488 hallmark of cancer development (Fasanelli et al. 2015).

489 Concurrent exposure to $PM_{2.5}$ at the time of blood sampling was associated with six hotspots
490 of differential expression and methylation in the present study. None of these appeared to
491 overlap with those identified in relation to $PM_{2.5}$ exposure at birth, which may at least partly
492 be attributed to differences in exposure levels as indicated by a moderate correlation between
493 birth and current exposure (0.32-0.46). It is also conceivable that exposure during different
494 time (age) periods of a growing child - from birth to adolescence will result in different
495 molecular responses. Among the six interactome hotspots we also found genes of importance
496 for lung adenocarcinoma, i.e. G protein subunit alpha i3 (*GNAI3*) (Ye et al. 2019), as well as

497 prostate cancer, *i.e.* *TAF8* (Alvarez and Woolf 2011). The finding of epigenetically
498 deregulated hotspot centred around the Scavenger receptor class A member 3 (*SCARA3*) gene
499 is of interest, in light of previous reports showing altered expression of this gene induced by
500 oxidative stress (Brown et al. 2013), one of the suggested key mechanisms involved in the
501 adverse health effects of air pollution. Further, differential methylation of this gene has been
502 linked to progression of type 2 diabetes mellitus (Karachanak-Yankova et al. 2016). Another
503 hotspot gene module, Tripartite motif containing 69 (*TRIM69*), has earlier been linked to the
504 incidence of heart failure (Karachanak-Yankova et al. 2016). It remains to be investigated
505 whether FEMs identified in this study can mediate the phenotypic health effects of air
506 pollution exposure on children's and young adults.

507 The present study is one of the first large-scale studies assessing the association of air
508 pollution exposure, represented by $PM_{2.5}$, during different periods of life on the children's
509 blood transcriptome. Coordinated gene expression measurements, quality control,
510 normalization procedures, as well as air pollution exposure assessment in all included studies
511 was performed according to a harmonized protocol. Further, integration of genome-wide
512 DNA methylation and gene expression data constitute another major strength of our study. All
513 cohort-specific analyses were conducted according to the same analytical protocol.

514 This study has some weaknesses. $PM_{2.5}$ exposure assessment was based on the participants'
515 residential address, without taking into consideration time-activity patterns and exposures at
516 nonresidential addresses (e.g. school). Further, the modeled concentrations account only for
517 outdoor air pollution and therefore may not be equivalent to the full range of personal
518 exposures. Another potential limitation is that we utilized purely spatial air pollution models
519 that were based on measurement campaigns carried out between 2008 - 2010 and applied
520 those to the home addresses of study participants over a period beginning from mid-1990s.
521 However, previous validation studies from Europe have demonstrated stability of spatial

522 contrast in levels of air pollution over time (Eeftens et al. 2011; Gulliver et al. 2011). Further,
523 we have explored this aspect in detail in our earlier investigations based on the same data by
524 performing sensitivity analyses using back-extrapolated modeled concentrations of considered
525 pollutants, and this had no impact on the results (Gehring et al. 2015; Gruzieva et al. 2014;
526 Molter et al. 2015). Although misclassification of air pollution exposure might have affected
527 our results, assessments of both exposure and gene expression levels were done independently
528 from each other, making such potential bias likely non-differential. It should also be noted
529 that we considered PM_{2.5} exposure at the birth address that may also be a proxy for prenatal
530 exposure. We, therefore, cannot rule out that at least part of the observed associations with
531 exposure at birth may be attributable to prenatal exposure. Furthermore, the study participants
532 were of European ancestry. Therefore, it is not sure if our findings can be generalized to other
533 ethnic groups. We should also acknowledge age differences between the cohorts included in
534 the present analyses and that our study was not designed to specifically evaluate potential age-
535 related effects in gene expression profiles. In addition, although we adjusted our analyses for
536 a number of potential confounders, still, the possibility of residual confounding cannot be
537 ruled out. Moreover, we used estimated cell counts to account for a potential cell type effect
538 in our analyses, as measured cell counts were not available in all cohorts. Finally, the omics
539 integration analysis assumes interaction at the protein level, whereas the networks probably
540 link to each other at the regulation level. Adding miRNA data in the integration analysis
541 would likely give more power, however the data is not available in the participating cohorts.
542 In conclusion, this study provides suggestive evidence for associations of PM_{2.5} exposure at
543 birth with differential gene expression in children and adolescents. Importantly, by integrating
544 gene expression and methylation data we could identify several significant interactome
545 hotspots of epigenetic deregulation gene modules in relation to PM_{2.5} exposure both at the
546 birth and at the current address. Our study shows the added value of integrating environmental

547 exposure data with multi-omics information to improve understanding of biologic responses

548 to exposure. Further studies are warranted to get deeper insight into the molecular

549 mechanisms for the harmful health effect of PM_{2.5}.

550

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